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Vertical integration of leaf area index in a Japanese deciduous broad-leaved forest

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Abstract

Leaf area index (LAI) is an important quantity in the study of forest ecosystems, but field measurements of LAI often contain errors because of the vertical complexity of the forest canopy. In this study, we established a practical method for field measurement of LAI in the canopy of a deciduous broadleaved forest by accounting for its vertical complexity. First, we produced a semi-empirical model for the vertical integration of leaf dry mass per unit leaf area. We also quantified the litterfall for each tree species. These data enabled us to estimate the LAI of each species in autumn. By periodic *in situ* monitoring of some fixed sample shoots throughout the growing season, we were able to estimate the seasonality of leaf area (as a proportion of the annual maximum value at each point in time) of each species. By using this seasonality to extrapolate LAI values as a proportion of the LAI data in the leaf-fall season, we were able to estimate LAI throughout the year. We applied this method in a cool-temperate deciduous forest in central Japan (Takayama) in 2006 and validated our results using two conventional methods of LAI measurement: the plant canopy analyzer (LAI-2000) and the Tracing Radiation and Architecture of Canopies (TRAC) approach. LAI estimated by TRAC was in good agreement with our results, but LAI estimated using the LAI-2000 was only half the value estimated using our method. The use of basal area data as a proxy for species-specific leaf areas may save labor and time. Our method will be useful for studying the dynamics and interactions of multiple species because it can estimate LAI and its seasonal changes for each species.

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37 Keywords: Leaf area index; Leaf mass per unit leaf area (LMA); Phenology; Litter

38 traps

1. Introduction

The leaf area in a forest canopy is an important quantity in understanding many ecosystem processes and characteristics of the forest. For example, photosynthesis, transpiration, aerodynamic roughness, autotrophic respiration, and other important parameters can be represented by integrating the functions of individual leaves for each tree's total leaf area. This idea, originating from the proposition by Monsi and Saeki (1953; republished in 2005), has been verified in various ecosystems from various standpoints (e.g., Hirose, 2005) and has become the theoretical basis for studying the functional interactions between a forest and its environment (e.g., Aber et al., 1996; de Pury and Farquhar, 1997; Watanabe et al., 2004; Walcroft et al., 2005). Most such studies have described leaf area using the leaf area index (LAI) parameter, which is defined as half of the total leaf area (i.e., the area for only one leaf surface) divided by the ground surface area beneath those leaves (e.g., Chen and Black, 1992; Jonckheere et al., 2004). However, field measurements of LAI often contain errors because of the vertical complexity of a forest canopy's structure, such as the mixture of multiple species with different characteristics, the presence of stems and branches, variation among the leaves (e.g., in their size, angle, thickness), and temporal changes in these factors.

Indirect optical methods for the field measurement of LAI, such as the LAI-2000 Plant Canopy Analyzer (Li-Cor Inc., Lincoln, NE, USA; e.g., Norman and Campbell, 1989), the Tracing Radiation and Architecture of Canopies approach (TRAC, 3rd Wave Engineering, Nepean, Ontario, Canada; Chen and Cihlar, 1995;

Leblanc and Chen, 2001), hemispherical photography (e.g., Muraoka and Koizumi, 2005) and light transmittance (e.g., Saigusa et al., 2002), assume a simple canopy architecture with little vertical complexity. Therefore, the accuracy of these methods is controversial (e.g., Jonckheere et al., 2004). In contrast, destructive sampling methods can directly provide the LAI value for each species (hereafter, the “component LAI”). However, it is difficult to monitor temporal changes in LAI with this method because we cannot always repeatedly destructively sample forest stands and because the method is prohibitively time-consuming. Allometric approaches (e.g., Macfarlane et al., 2007) can replace destructive sampling, but it remains difficult to monitor seasonal changes using this approach.

For a deciduous forest, it is possible to overcome some of these problems using litter traps (e.g., Jonckheere et al., 2004): If leaf expansion and leaf fall do not take place simultaneously, we can estimate the biomass of leaves in the canopy during the leaf-fall season by collecting leaf litter in litter traps on the forest floor. We can then estimate the component LAI by sorting the leaf litter by species (e.g., Tateno et al., 2005). However, this method provides little information about LAI during the leaf-expansion season. Moreover, to convert the biomass value into leaf area values, we must calculate either the leaf dry mass per unit of leaf area (LMA) or the specific leaf area (SLA), which is the reciprocal of LMA. In order to estimate LMA or SLA, we must account for their vertical variability (e.g., Eriksson et al., 2005).

In the present study, we aimed to develop a practical method for field

measurement of LAI in the canopy of a deciduous broadleaved forest that accounts for the canopy's vertical complexity and for seasonal changes. Our method combines three components. First, we propose a semi-empirical model for the vertical integration of LMA within a canopy. Second, using this model and litter-trap observations, we estimate the component LAI during the leaf-fall season. Third, by periodically obtaining *in situ* observations of sample shoots, we can estimate the pattern of relative change (hereafter, "seasonality") of the component LAIs. By scaling the seasonality of the component LAIs so that they are continuously connected to the component LAI in the leaf-fall season, we can extend our estimation of component LAI into the leaf-expansion season. By adding the component LAI of all species, we can then obtain LAI for the entire canopy from the initial leaf expansion to leaf fall. We implemented this method in a Japanese cool-temperate deciduous broadleaved forest. Using this result as a reference value, we validated it against two conventional indirect optical methods (LAI-2000 and TRAC).

2. Materials and methods

2-1. Study site

The study site is a cool-temperate deciduous broadleaved forest near Takayama, in central Japan (137.4231°N, 36.1462°E, 1420 m a.s.l.; Fig. 1). The annual mean air temperature and the mean annual rainfall from 1980 to 2002 were 7.2°C and 2275 mm, respectively. The site is covered by a snowpack that ranges from 100 to 180 cm

in depth from December until April (Mo et al., 2005). A tree census has been carried out every year since 1999 (Ohtsuka et al., 2005). The fluxes of CO₂, water vapor, and sensible heat have been observed at a flux tower since 1993 (Yamamoto et al. 1999; Saigusa et al. 2002, 2005; Fig. 1). Leaf physiology (photosynthesis and respiration) of some dominant species has been observed using a canopy-access tower that is 18 m tall (Muraoka and Koizumi, 2005; Fig. 1). Some ecophysiological models have been developed and tested for this site (Higuchi et al., 2005; Ito et al., 2005, 2006).

(Insert Fig. 1 here)

The species composition of the canopy trees at the study site is as follows: The dominant tree group consists of *Quercus crispula* and *Betula ermanii*, with some *Fagus crenata*, *Betula platyphylla*, and *Magnolia obovata*. The co-dominant trees include *Acer distylum*, *Acer rufinerve*, *Acanthopanax sciadophylloides*, *Tilia japonica*, *Sorbus alnifolia*, and *Kalopanax pictus*. The suppressed trees include *Hydrangea paniculata* and *Viburnum furcatum*. Under the suppressed trees, the forest floor is covered by an understory of evergreen dwarf bamboo (*Sasa senanensis*) with a height of 1.0 to 1.5 m. The height of the dominant forest canopy ranges from 13 to 20 m.

Most observations were carried out from April to November 2006. For the study area, the leaf-expansion season is the period from April to mid-August, and the leaf-fall season is the period from late August to November.

2-2. A model for the vertical integration of LMA within a canopy

Even within the same species or the same individual, leaf characteristics vary within the canopy, mainly due to variations in light conditions. For the sake of simplicity, we have considered only two extreme cases of this variation as a function of height, namely the leaf at the top position on the stem (hereafter, the “top leaf”) and a leaf at the lowest position (hereafter, called “bottom leaf”). Using these two extremes, we can approximate the entire canopy as a mixture of these two leaf types.

In our model, we define the z -axis as extending vertically downward from the canopy surface ($z = 0$). At the forest floor, $z = Z$. We denote total LAI as L . We denote the cumulative leaf area over a unit area of ground from $z = 0$ to z as $l(z)$. Obviously, $l(0) = 0$ and $l(Z) = L$. Let $m = \text{LMA}$, which depends on z .

To begin our analysis, let us consider a canopy composed of a single species. For a thin layer between z and $z + dz$ within the canopy, the leaf area in this layer (dL) is $dL = L(z + dz) - L(z)$. The leaf dry mass (dM) in this layer is represented as: $dM = m dL$. Therefore, the total dry mass (M) of the entire canopy per unit of ground area is:

$$M = \int dM = \int_0^L m dl . \quad (1)$$

Niinemets and Tenhunen (1997) proposed the following relationship based on field study data and a semi-empirical model of a deciduous broadleaved forest:

$$m = a \exp (-bL), \quad (2)$$

where a and b are constants that depend on the species and site. Dufrêne et al. (2005) adopted this function in their numerical model. We have assumed that this formula is generally valid for deciduous broadleaved forests.

We denote the LMA values of the top leaf and the bottom leaf as m_0 and m_1 , respectively. From equation (2), it is obvious that $m_0 = m(0) = a$ and $m_1 = m(Z) = m_0 \exp(-bL)$. Therefore,

$$a=m_0 \quad \text{and} \quad b = \frac{1}{L} \ln \frac{m_0}{m_1}. \quad (3)$$

By substituting equations (2) and (3) into equation (1), we get:

$$M = L \frac{m_0 - m_1}{\ln m_0 - \ln m_1}. \quad (4)$$

From this equation, the average LMA (m_a) of the entire canopy can be defined as the total dry leaf mass divided by the total leaf area:

$$m_a = \frac{M}{L} = \frac{m_1 - m_2}{\ln m_1 - \ln m_2} . \quad (5)$$

Now, let us extend this formula to a complex canopy with multiple species.

First, let us consider a layer occupied by a single species between $z = z_2$ and $z = z_3$ in the canopy. We describe $l(z_2)$ as l_2 and $l(z_3)$ as l_3 . We can describe LMA at z_2 and z_3 as m_2 and m_3 , respectively. We assume that equation (2) is valid for the layer occupied by the single species within the complex canopy. From this assumption, $m_2 = a \exp(-bl_2)$ and $m_3 = a \exp(-bl_3)$. Therefore,

$$b = \frac{1}{L_{23}} \ln \frac{m_2}{m_3} . \quad (6)$$

where, $L_{23} = l_3 - l_2$. The leaf dry mass in this layer (M_{23}) can be calculated in the same way as equation (4):

$$M_{23} = L_{23} \frac{m_2 - m_3}{\ln m_2 - \ln m_3} \quad (7)$$

If we define the average LMA in this layer (m_{a23}) as M_{23} / L_{23} ,

$$m_{a23} = \frac{M_{23}}{L_{23}} = \frac{m_2 - m_3}{\ln m_2 - \ln m_3} , \quad (8)$$

which is similar to equation (5).

Now let us consider how the canopy (or a layer) of a single species can be represented as a mixture of top and bottom leaves. Because the LAI of the canopy is the sum of the LAI of the two types of leaves, $L = L_0 + L_1$. Then, the total leaf dry mass (M) can be represented as $M = m_0 L_0 + m_1 L_1$. This must be equal to equation (4). Therefore,

$$m_0 L_0 + m_1 L_1 = L \frac{m_0 - m_1}{\ln m_0 - \ln m_1} . \quad (9)$$

We can simplify this equation to eliminate L_1 by using the relationship $L = L_0 + L_1$, producing the following formula:

$$\frac{L_0}{L} = \frac{1}{\ln m_0 - \ln m_1} - \frac{m_1}{m_0 - m_1} . \quad (10)$$

This tells us about the fraction of top leaves in the canopy. We can estimate the fraction of the bottom leaves by subtracting this result from 1.

In a field study in southern Swedish forest stands, Eriksson et al. (2005) used equal proportions of the top and bottom leaves when estimating average tree SLA. In that case, the equation corresponding to equation (8) is $1 / m_{ae} = [(1 / m_0) + (1 / m_1)] / 2$, or equivalently:

$$m_{ae} = \frac{2 m_0 m_1}{m_0 + m_1} \quad (11)$$

where m_{ae} is the average LMA under the assumption of Eriksson et al. (2005).

2-3. Observation of LMA

In order to determine the LMA of each species, we sampled the top and bottom leaves of each dominant species in August, September, and October in 2006 and 2007. We measured LMA of each leaf as follows: First, we measured the area of each leaf by scanning the leaf with an ES-7000H image scanner (SEIKO EPSON Co., Japan, 300 dpi resolution). We calculated the leaf's area by counting the number of leaf pixels and multiplying this total by the pixel size using the LIA-32 image-analysis software (<http://www.agr.nagoya-u.ac.jp/%7Eshinkan/LIA32>). Most leaves were flat. The non-flat leaves (wrinkled or rolled) were flattened by the lid of the image scanner. After the area measurement, we dried the leaf samples in an oven at 80°C for more than 48 h. We then measured the dry mass of each sample. By dividing the dry mass by the leaf area, we obtained LMA for each sample. Then, using equation (8), we obtained the average LMA for each species. Strictly speaking, LMA can change along with growth and aging of the leaves. However, we neglected this change because the change was estimated to be insignificant so far as we limit it in the late summer and autumn.

2-4. Litter-trap observations

At the study site, we installed 14 litter traps within a 1-ha permanent sample plot that is being used for ecological surveys (Fig. 1). Each trap had a square aperture of 1 m². Their locations were chosen so that they covered a wide range of site topographic conditions (ridge, slope, and valley) and tree species. In particular, we avoided placing multiple traps under the crown of a single tree. The litter (leaves, branches, seeds, etc.) caught in the traps were recovered on August 25, September 17, September 30, October 9, October 22, November 4, and November 18. During each litter collection, we sorted the litter in each trap into the leaves of each species. Because it was difficult to distinguish between the leaves of *B. ermanii* and *B. platyphylla*, we treated them as a single *Betula* category. The tree census data for this site (Ohtsuka et al., 2005) suggest that *B. ermanii* occupied 63% of the total basal area in the *Betula* category. We discarded the leaves of evergreen trees such as *Pinus parviflora* because their amount was small. In fact, the evergreen trees occupied only 2.8% of the site's basal area and only 1.0% of the individuals that formed the entire canopy (Ohtsuka et al., 2005). After oven-drying the leaves at approximately 70°C for longer than 48 h, we measured the mass of the dried leaves. By dividing this mass by the average LMA, we estimated the area of the fallen leaves of each species at the time of sampling. By adding the area of the fallen leaves from August 25 to November 18, we obtained a total LAI, and were able to estimate the component LAI at each point between these dates for the leaf-fall season.

2-5. Leaf seasonality observations

We carried out leaf seasonality observations by means of periodic *in situ* observation of sample shoots. We selected 20 shoots of 18 individuals of 8 species for these samples, all of which we could directly access from the canopy access tower or the forest floor (Table 1). The selection of the 8 species was based on their relative rank in the total tree biomass for the site (Ohtsuka et al., 2005). The selection of the 20 sample shoots was based on ease of access and the goal of measuring shoots at a variety of positions.

(Insert Table 1 here)

On May 12, May 22, June 2, June 16, June 29, July 30, August 18, September 17, September 30, October 9, October 22, and November 5, we obtained the following observations for each sample shoot: the number of all leaves on the shoot and the size (length and width) of about 20 randomly selected leaves on each shoot. By approximating the shape of the leaves as an ellipse and assuming that the measured length and width represented the longest and shortest axes of the ellipse, respectively, we estimated the area of each leaf. Summation of these leaf areas gave us an estimate of total leaf area on the shoot at that point in time. Thus, we obtained a time series for the seasonal changes in total leaf area on each shoot. By normalizing these data so that the seasonal peak value became 1.0, we obtained the seasonality of the leaf area (i.e., the proportion of the maximum value at each point in time) for each shoot. By averaging these results within each species, we obtained the seasonality of the overall leaf area of each species. For *Betula* and *Quercus*, we

averaged the seasonality at the top position and at the bottom position separately,
and then averaged the result using a fractional weight derived from equation (10).

These seasonality data were assumed to represent the seasonality of the
component LAI. The average seasonality of *A. distylum* and *A. rufinerve* was
assumed to represent the seasonality of the other codominant tree species. In the
same way, the average seasonality of *H. paniculata* and *V. furcatum* was assumed to
represent the seasonality of other suppressed tree species.

2-6. Development of seasonal LAI

We multiplied the seasonality of the component LAI so that it continuously
connected to the component LAI for the leaf-fall season estimated by the litter-traps.
This allowed us to extrapolate our results to obtain estimates of component LAI
throughout the leaf-expansion season. By combining these results, we were able to
estimate component LAI throughout the growing season. We also obtained the total
LAI of the entire canopy (excluding *Sasa*) by adding the component LAI values for
all the species.

LAI of *Sasa* on the forest floor in this site was estimated in April 1999 at 119
sampling points (Sakai et al., 2002), and the result was a mean of 1.71, with a
standard deviation of 0.933. Assuming statistical independence among the 119
samples, the standard error of the mean was 0.09. Leaf biomass of *Sasa* at this site
was mostly stable for all seasons in 1993 and 1994 (Nishimura et al., 2004). Based
on these two results, we assumed a constant LAI of 1.71 for *Sasa*, with a standard

error of 0.09.

2-7. Observations of LAI using indirect optical methods

We used LAI-2000 to measure LAI on May 14, June 2, June 30, July 31, September 8, and October 23. We used two LAI-2000 instruments simultaneously: one to measure incoming light from the sky above the canopy, and the other to measure transmitted light at ground level. The LAI-2000 at the top of the canopy access tower was operated in automatic mode, with measurements recorded at 15-s intervals. The LAI-2000 at ground level was moved around the site by a researcher to obtain measurements above the *Sasa* canopy (approximately 1.5 m above the ground), taking five measurements within 10 s each time at each location (mostly at or near the litter traps). LAI-2000 measurements were also obtained below the *Sasa* canopy (approximately 0.1 m above the ground) on May 14 in order to estimate the LAI of *Sasa*. We analyzed the data from the two LAI-2000s using the Li-Cor FV2000 software (version 1.06), with all the initial parameters and conditions left at the software's default values.

We performed TRAC measurements on June 3 and June 29 along the two 100-m transects shown in Figure 1. During each measurement, we also monitored the photosynthetic photon flux density (PPFD) every 5 s with a quantum sensor (IKS-27, Koito Industries, LTD., Yokohama, Japan) at the top of the canopy access tower. We analyzed the TRAC data using the TRACWIN software (version 3.9.1). We tested four groups of parameter settings for each TRAC measurement using two

values for “PPFD above” (the maximum and minimum PPFD values measured at the canopy tower) and two values for “mean element width” (the upper and lower limits of the range within one standard error of the mean leaf diameter estimated using the litter-trap and leaf-seasonality observations).

In general, these indirect methods inevitably observe not only leaves but also stems and branches within their field of view. Therefore, their results are not true LAI values, but rather PAI (plant area index) values, which equal the sum of LAI and SAI (stem area index; the total stem and branch area per unit of ground surface area). We assumed SAI to be invariant and equal to the PAI value observed with the LAI-2000 on May 14, when all the deciduous trees had no leaves. By subtracting this SAI value from the PAI values taken by the LAI-2000 and TRAC, we estimated LAI. Because we obtained these indirect observations above *Sasa* on the forest floor in most cases, these PAI and LAI values excluded the values for *Sasa*.

3. Results

3-1. LMA

We obtained LMA values for the study species that ranged from 25 to 92 g m⁻² (Table 2). The standard error of each LMA value was less than 6% of the mean. In all species except for *V. furcatum*, the LMA of the top leaves was 1.5 to 2.1 times that of the bottom leaves; for *V. furcatum*, there was little difference between the two leaf types. The average LMA derived using equation (8) ranged from 32 to 76 g m⁻².

(Insert Table 2 here)

We tested equation (11) with our data and found that the resulting average LMA was 74.4 g m⁻² for *B. ermanii*, 68.5 g m⁻² for *Q. crispula*, and 30.7 g m⁻² for *H. paniculata*. These values were smaller than the average LMA calculated using equation (8), but the difference was less than the standard error.

3-2. Litter-trap observations

As of late summer (August 25), the canopy LAI (excluding *Sasa*) equaled 5.0, of which 70% was accounted for by the dominant group, 16% by the codominant group, and 14% by the suppressed group. The standard error of the total LAI was 0.3. This error was attributable to both the limited number of the litter traps (14) and the error associated with LMA.

3-3. Leaf-seasonality observations

All the species showed clear seasonality of leaf number, single-leaf size, and leaf area per shoot (Fig. 2). As seen in the changing number of leaves (left column of Fig. 2), most species except *B. ermanii* had a single leaf flush (a rapid emergence of leaves) in May. In contrast, *B. ermanii* showed two leaf flushes: the first one in late May and the second in late June. Because of the small new leaves produced during the second flush, the mean single-leaf area of *B. ermanii* decreased in late June. The new leaves in the second flush never grew as large as the leaves in the first flush, thus the mean single-leaf area after June did not recover to the maximum level

attained in May. For some species (*Q. crispula* and *F. crenata* in particular), the single-leaf area decreased slightly from the leaf flush until August, probably because of herbivory by insects.

(Insert Fig. 2 here)

After summer, the single-leaf area of some species showed irregular changes. For example, the single leaf area of *M. obovata* decreased temporarily in early October and then recovered. This irregularity could occur in the following manner: when large leaves on the shoot fell, the average single-leaf area decreased temporarily, but thereafter, when small leaves fell, the average single-leaf area recovered. Conversely, the single-leaf area of *F. crenata* temporarily increased in October. Such irregularity is partly attributable to statistical errors resulting from the small sample size.

The seasonality of leaf area per shoot during the leaf-fall season differed even within a species (particularly for *B. ermanii* and *H. paniculata*), so that the standard error of the leaf area was larger than that during the leaf-expansion season. Except for the leaf-fall period, the standard error of seasonality of leaf area per shoot was mostly less than 0.1. This magnitude of error suggests that the selection of 20 sample shoots offers a valid method for representing the seasonality of the component LAI only during the leaf-expansion season.

3-4. LAI in all seasons

By combining the litter-trap data with the leaf seasonality data, we were able to

estimate the component LAI during all seasons (Fig. 3). *Betula* had the largest peak LAI, followed by *Q. crispula*.

(Insert Fig. 3 here)

The total LAI (Fig. 4) also showed clear seasonal changes with a maximum of 7.0 on July 30. Throughout most of the study period, the dominant group accounted for the majority of the total LAI. However, *Sasa* accounted for the majority of LAI in May and November because most of the deciduous trees had no leaves on their branches at this time. The codominant group and the suppressed group also accounted for higher fractions than the dominant group in mid-May because their leaf flush happened earlier than that of all trees in the dominant group except for *F. crenata*.

(Insert Fig. 4 here)

3-5. LAI estimation using indirect optical methods

SAI was estimated to be 0.8 from the LAI-2000 observations on May 14. By subtracting this SAI value from PAI, we estimated LAI with the LAI-2000 and the TRAC approach (Fig. 5). The LAI-2000 gave obviously lower LAI values (by about half) than those provided by our method throughout the study. Moreover, the pattern of seasonal change was different: in July, the LAI-2000 showed a continuing increase of LAI (by as much as 20% of the annual peak), whereas our method showed little increase. In contrast, the TRAC LAI values were close to those provided by our method. The LAI of *Sasa* was estimated to be 1.55 (with a standard

error of 0.10, $n = 21$) using the LAI-2000 observations on May 14, which was close to our assumption of 1.71.

(Insert Fig. 5 here)

4. Discussion

4-1. Reliability and utility of the proposed method

The average LMA of each species derived using equation (8) showed little difference from the average LMA calculated using equation (11). However, if L_0 and L_1 differ greatly, the choice of equation becomes more critical. In order to address this issue, we need further tests of the relationship between the cumulative leaf area and the LMA proposed by Niinemets and Tenhunen (1997). This would require stratification of the number of the fallen leaves (leaf litter) in each LMA category, something we could not do in the present study because it was difficult to measure the area of the fallen leaves, which were deformed and distorted as a result of drying between collection dates.

The eight species that we selected for the leaf-seasonality observations accounted for 84% of LAI (on August 25, excluding *Sasa*) estimated from the litter-trap data. This means that the seasonality of about 16% of total LAI was uncertain. Therefore, in order to attain more accurate assessments of the seasonality of LAI, we should obtain observations for more species (especially for *T. japonica*) that were ignored in the present study.

From a practical standpoint, our method offers the advantage that it can

work under any weather and light conditions. In contrast, the LAI-2000 and TRAC approaches require suitable weather and light conditions, which are not always available. However, our method is laborious in comparison to these alternatives. In particular, the process of sorting the leaf litter requires considerable time and a certain amount of expertise. If we could eliminate this part of the process, our method would become more convenient. To do so, it would be helpful to have a reliable proxy for the component LAI that is easier to measure in the field.

Figure 6 shows that the fraction of the component LAI (excluding evergreen trees and *Sasa*) of the major species in August corresponded well to the fraction of their total basal area estimated from a tree census study (Ohtsuka et al., 2005). This relationship is consistent with a study of the allometric relationship for 46 North American deciduous species, in which Niklas (1994) found that leaf area increased with the square of stem diameter (i.e., with increasing basal area). Based on this relationship, it may be possible to eliminate the process of sorting the leaf litter.

(Insert Figure 6 here)

4-2. Validation of LMA used in previous studies

From the total leaf dry mass of 315 g m^{-2} and the total LAI (excluding *Sasa*) of 5.0 on August 25, the LMA for the entire canopy (excluding *Sasa*) was estimated as 63 g m^{-2} . In comparison, a previous ecophysiological process model (Ito et al., 2005) reported LMA equal to 67 g m^{-2} (based on an SLA value of $150 \text{ cm}^2 \text{ g}^{-1}$), which is slightly higher than our value. On the other hand, Ito et al. (2006) reported that

LMA equaled 93.1 g m⁻² for *Q. crispula* and 69.2 g m⁻² for *Betula* in another modeling study, and these values differed greatly from our estimates of 69.7 and 76.3 g m⁻², respectively (Table 2). We believe that this discrepancy is mostly attributable to the vertical scaling that we performed, which was not attempted in previous studies at this site.

4-3. Comparison with indirect optical methods of LAI estimation

Obviously, it is impossible for every study to carry out such detailed measurements as those in the present study. Therefore, it is desirable to validate and improve the indirect optical methods because they are easier and quicker to carry out. In comparison with our method, the TRAC method provided comparable LAI values (although we did not validate this approach for the later parts of the growing season), but the LAI-2000 gave lower LAI values and a dissimilar pattern of seasonal change. One possible cause for this discrepancy is the clumping effect (e.g., Leblanc and Chen, 2001), which is accounted for by the TRAC method but not by the LAI-2000 method. The clumping index must equal about 0.5 to explain the discrepancy between these two instruments. However, the actual clumping index estimated using the TRAC data ranged between 0.91 and 0.95. Therefore, the clumping index alone cannot explain the discrepancy.

Our results also suggest underestimation of LAI values derived by indirect methods in previous studies at this site. In 1997 and 1998, Saigusa et al. (2002) used transmittance of PPFD to estimate the annual peak PAI (excluding *Sasa*),

which ranged between 3 and 4. In 2003, using hemispherical photography, Muraoka and Koizumi (2005) estimated an annual peak PAI of 4.0 (excluding *Sasa*) at 14 locations at this site (mostly corresponding to the litter traps in the present study). If we assume that $SAI = 0.8$, then the annual peak LAI (excluding *Sasa*) in these two studies should be between 2.2 and 3.2, which is smaller than our estimate of 5.3 on July 30 for the annual peak LAI (excluding *Sasa*). However, the PAI values reported by Muraoka and Koizumi (2005) showed seasonal changes that were consistent with the pattern revealed using our proposed method.

To learn the reasons for the discrepancy among the methods, we will need to validate each step in the derivation of LAI in each method, which is the next step after the present study. Along with the leaf seasonality observations, we also observed the distribution of leaf angles and leaf transmittance of each tree species. These data, which we will present in a future paper, should provide more detailed evidence for the causes of errors in the indirect methods.

4-4. Ecological implications

In general, the tree species in cool-temperate deciduous forests in Japan are categorized into “single leaf flush” types such as *Fagus* spp., *Quercus* spp., and *Acer* spp., and “indeterminate leaf production” types such as *Betula* spp. (Koike, 1988; Kikuzawa, 2005). Trees of the “single leaf flush” type have a leaf flush that occurs within a single short period in the early growing season, whereas trees of the “indeterminate leaf production” type have either a gradual leaf flush or several

consecutive leaf flushes. The former type has smaller LMA values than the latter type. All these characteristics are considered to be related to their adaptation strategy to maximize light acquisition and utilization. In this study, these characteristics were obvious (Table 2, Fig. 2, Fig. 3).

These characteristics were reflected in the seasonal changes in component LAI. For example, trees in the codominant group accounted for the majority of total LAI (excluding *Sasa*; Fig. 4) early in the growing season (May). This may relate to the favorable light conditions created by the slower leaf expansion of the dominant trees. Such favorable light conditions for the codominant group, which has been called the “seasonal gap” (Yamamoto, 2000), should depend on the species composition and ecophysiological characteristics of the dominant group. For example, if *Q. crispula* become the majority of the dominant group (currently *Betula* spp.), the earlier leaf-flush and later leaf-fall of *Q. crispula* will create a shorter “seasonal gap,” resulting in a more severe environment for the trees below this canopy. Such an influence could be demonstrated using a numerical model that describes the regeneration process as well as the seasonality of leaf area for each species. This modeling would give insights into the consequences of interactions between forest canopy structure, the resulting light environment, and regeneration processes (Pearcy, 1990; Kupperts, 1994), all of which are affected by the vertical complexity of the canopy. Our method can contribute to such quantitative studies by providing fundamental data that can be used to explain the seasonal changes in the component LAI.

5. Conclusions

We were able to estimate the total LAI of a Japanese deciduous broadleaved forest by accounting for the vertical complexity of the canopy and its seasonal changes. This method can provide not only total canopy LAI but also the component LAI (i.e., LAI values for each species) and its seasonal changes. In this method, it may be possible to eliminate the most laborious step (sorting of leaf litter by species) if the basal area of each species is available. Our method provided considerably higher LAI values than those estimated using the LAI-2000 approach, but comparable values to those provided by the TRAC approach.

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Captions

Table 1. Sample shoots for the *in-situ* leaf-seasonality observation

Table 2. Leaf mass per unit leaf area (LMA; gm^{-2}) of the major tree species

Fig. 1. Top: Location of the Takayama (TKY) site in Japan. Bottom: Arrangement of the litter traps and the Tracing Radiation and Architecture of Canopies (TRAC) measurement transects at the Takayama Site. Each litter trap is identified with a label made of a symbol “L” and a two-digit number from L03 to L17 (L01, L02, and L06 do not exist).

Fig. 2. Seasonality of leaf characteristics for the tree species. For *B. ermanii* and *Q. crispula*, the top and bottom shoots in the canopy are presented separately. Error bars represent the standard error. Each time series for the data was normalized using the annual maximum value set to 1.0.

Fig. 3. Leaf area index (LAI) of the tree species (component LAI) estimated by combining the leaf-seasonality data (until August 25) with the litter-trap data (after August 25).

Fig. 4. Top graph: Leaf area index (LAI) of the entire canopy and of the dominant, codominant, and suppressed tree groups, and the understory layer (*Sasa* dwarf bamboo), estimated by the proposed method. Bottom graph: The corresponding fractions of total LAI for each group.

Fig. 5. Leaf area index (LAI) of the canopy (excluding *Sasa*) estimated by three methods: the method proposed in this study (which combines litter-trap data with leaf-seasonality observations), the LAI-2000 canopy analyzer, and the Tracing

Radiation and Architecture of Canopies (TRAC) approach. The LAI values for the LAI-2000 and TRAC approaches were estimated from the observed plant area index (PAI) under the assumption that the stem area index (SAI) equaled 0.8. The error bar represents standard error except for TRAC. The error bar of TRAC represents the range between the minimum and maximum of the LAI values calculated with the four groups of the parameter setting.

Fig. 6. Fraction of total LAI accounted for by the major species at the study site (*Betula*, *B. ermanii* and *B. platyphylla* combined; *Quercus crispula*; *Magnolia obovata*; *Acer rufinerve*; *Tilia japonica*; *Hydrangea paniculata*; and *Viburnum furcatum*) in late summer (August 25) versus the corresponding fraction of total basal area (Ohtsuka et al., 2005). Error bars represent the standard error.

Table 1. Sample shoots for the *in-situ* leaf-seasonality observation

		ID of the individual	Height of the shoot	Leaf positions	Maximum number of leaves
	Dominant tree group				
	<i>Betula ermanii</i>	Be_A	14 m	top	66
		Be_B	18 m	top	51
			15 m	bottom	
165					
		Be_C	16 m	bottom	55
	<i>Quercus crispula</i>	Qc_A	15 m	top	94
		Qc_B	14 m	top	149
			10 m	bottom	80
		Qc_C	11 m	bottom	93
		Qc_D	14 m	top	50
	<i>Magnolia obovata</i>	Mo_A	14 m	middle	34
	<i>Fagus crenata</i>	Fc_A	1.2 m	bottom	59
	Codominant tree group				
	<i>Acer distylum</i>	Ad_B	1.3 m	bottom	37
		Ad_C	1.3 m	bottom	42
	<i>Acer rufinerve</i>	Ar_A	4.0 m	middle	72
	Suppressed tree group				
	<i>Hydrangea paniculata</i>	Hp_A	1.5 m	middle	163
		Hp_B	0.8 m	middle	68
		Hp_C	1.5 m	middle	96

697	<i>Viburnum furcatum</i>	Vf_A	2.5 m	middle	91
698		Vf_B	1.0 m	middle	18
699		Vf_C	1.0 m	middle	20

700

Table 2. Leaf mass per unit leaf area (LMA; g m⁻²) of the major tree species

Species	Position	Mean	s.d.	Sample size	std. error
Dominant group					
<i>Betula ermanii</i>	top	92.2	4.3	5 leaves	2.1
	bottom	62.3	3.6	5 leaves	1.8
	average	76.3	---	---	2.8
<i>Quercus crispula</i>	top	81.9	15.7	13 leaves	4.5
	bottom	58.8	10.0	8 leaves	3.8
	average	69.7	---	---	5.9
<i>Magnolia obovata</i>	top	91.6	0.1	2 leaves	0.1
	bottom	44.0	0.2	2 leaves	0.2
	average	64.9	---	---	0.2
<i>Fagus crenata</i>	top	79.9	6.1	5 leaves	1.6
	bottom	52.0	3.5	2 leaves	4.3
	average	64.9	---	---	2.6
(Mean of averages)		69.0	---	---	5.6
Codominant group					
<i>Acer distylum</i>	top	44.0	2.3	5 leaves	1.2
	bottom	24.6	2.3	5 leaves	1.2
	average	33.4	---	---	1.7
<i>Acer rufinerve</i>	top	64.0	6.8	5 leaves	3.4
	bottom	31.3	2.5	5 leaves	1.3
	average	45.7	---	---	3.6
<i>Acanthopanax</i>	top	42.5	1.3	5 leaves	0.7

725	<i>sciadophylloides</i>	bottom	27.7	0.4	5 leaves	0.2
726		average	34.6	---	---	0.6
727	(Mean of averages)		37.9 ¹⁾	---	---	6.9
728	(Mean of averages, <i>Acer</i> only)		39.5 ²⁾	---	---	9.1
729	Suppressed group					
730	<i>Hydrangea paniculata</i>	top	39.9	3.8	5 leaves	1.9
731		bottom	25.0	2.0	5 leaves	1.0
732		average	31.9	---	---	2.1
733	<i>Viburnum furcatum</i>	top	43.3	1.9	5 leaves	0.9
734		bottom	43.5	4.3	5 leaves	2.1
735		average	43.4	---	---	2.3
736	(Mean of averages)		37.6 ³⁾	---	---	8.5

737 Note:

738 “top” means leaf samples from the highest part of the tree crown; “bottom” means the leaf

739 samples from the lowest part of the tree crown; “s.d.” means standard deviation; “std. error”

740 means standard error (standard deviation of the mean). The boldfaced “**average**” values means

741 the averaged LMA based on equation (8), whereas the “mean of averages” means the

742 arithmetic mean of the **averaged** LMA.

743 1) This value was used as the surrogate LMA for all other codominant species, including *T.*

744 *japonica*, *S. alnifolia*, and *K. pictus*.

745 2) This value was used as the surrogate LMA for all other *Acer* species.

746 3) This value was used as the surrogate LMA for all other suppressed species.

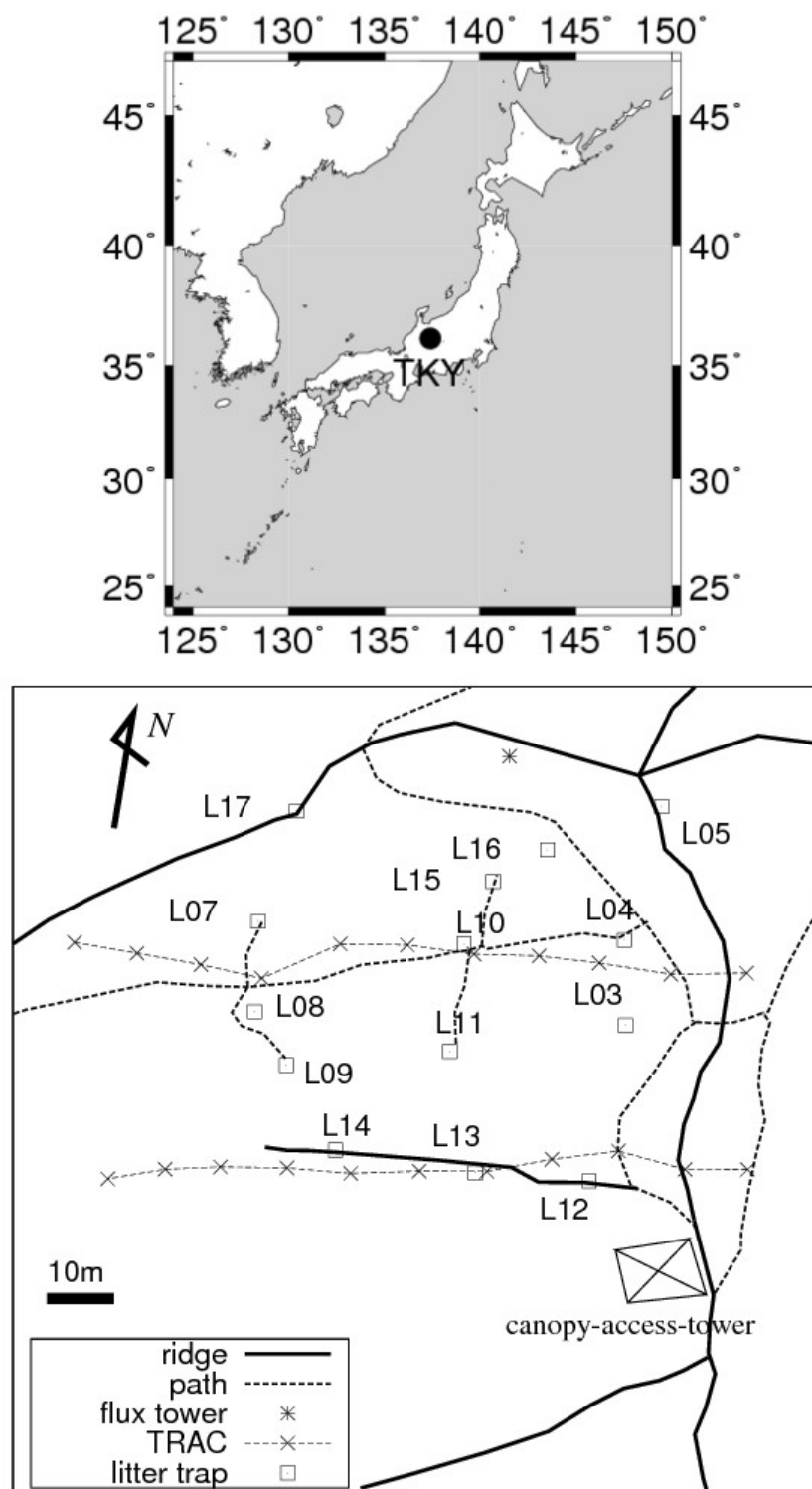


Fig. 1

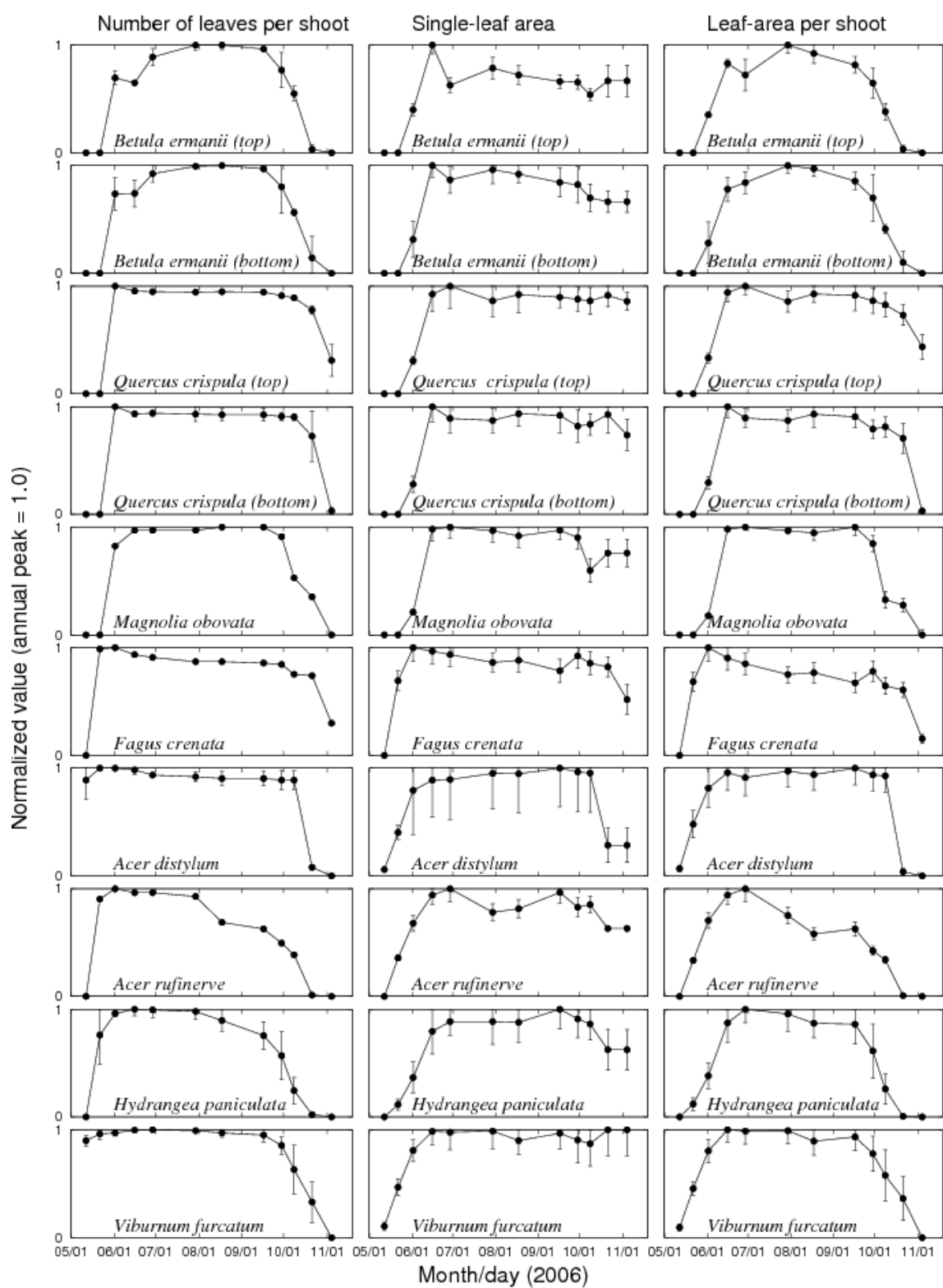


Fig. 2

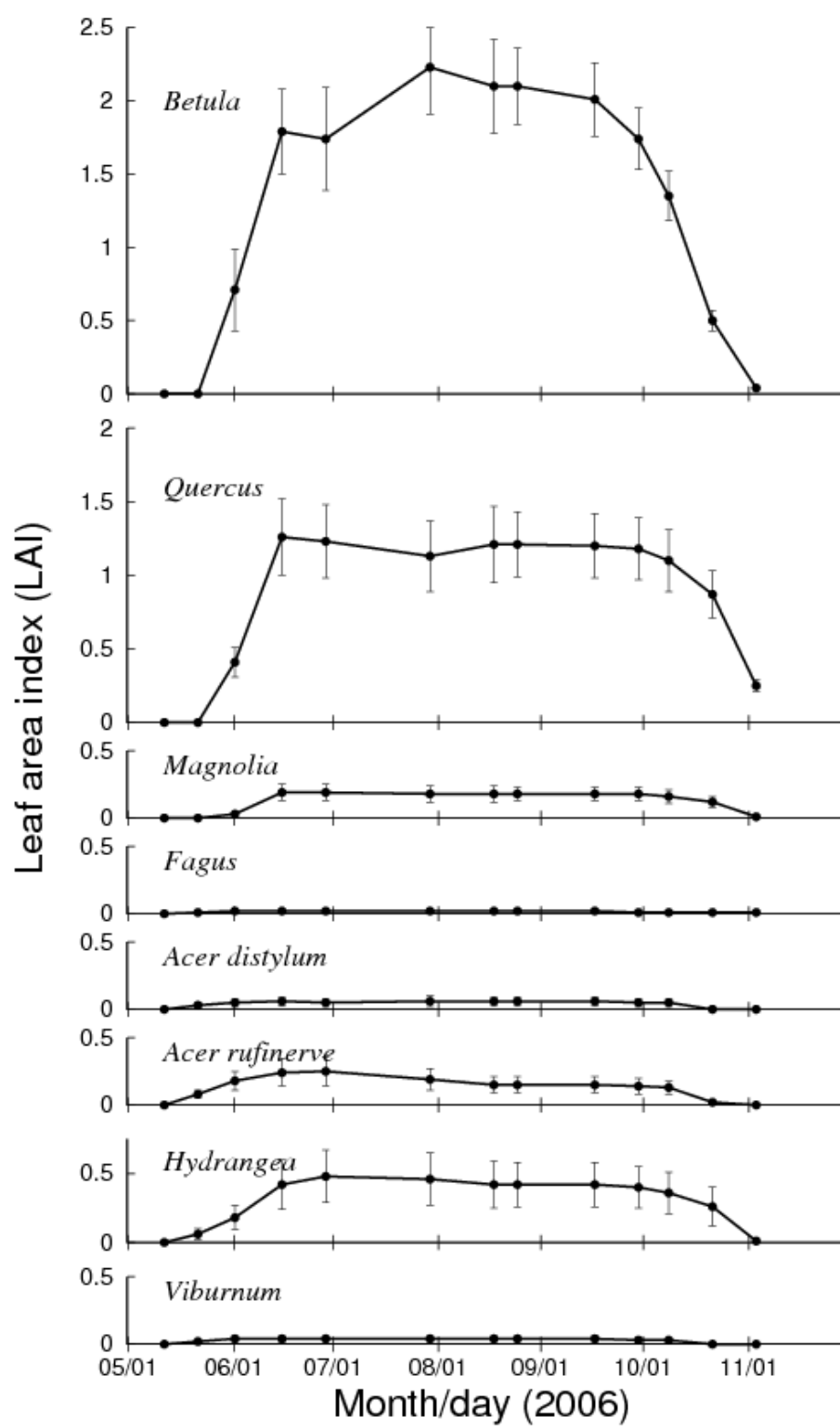
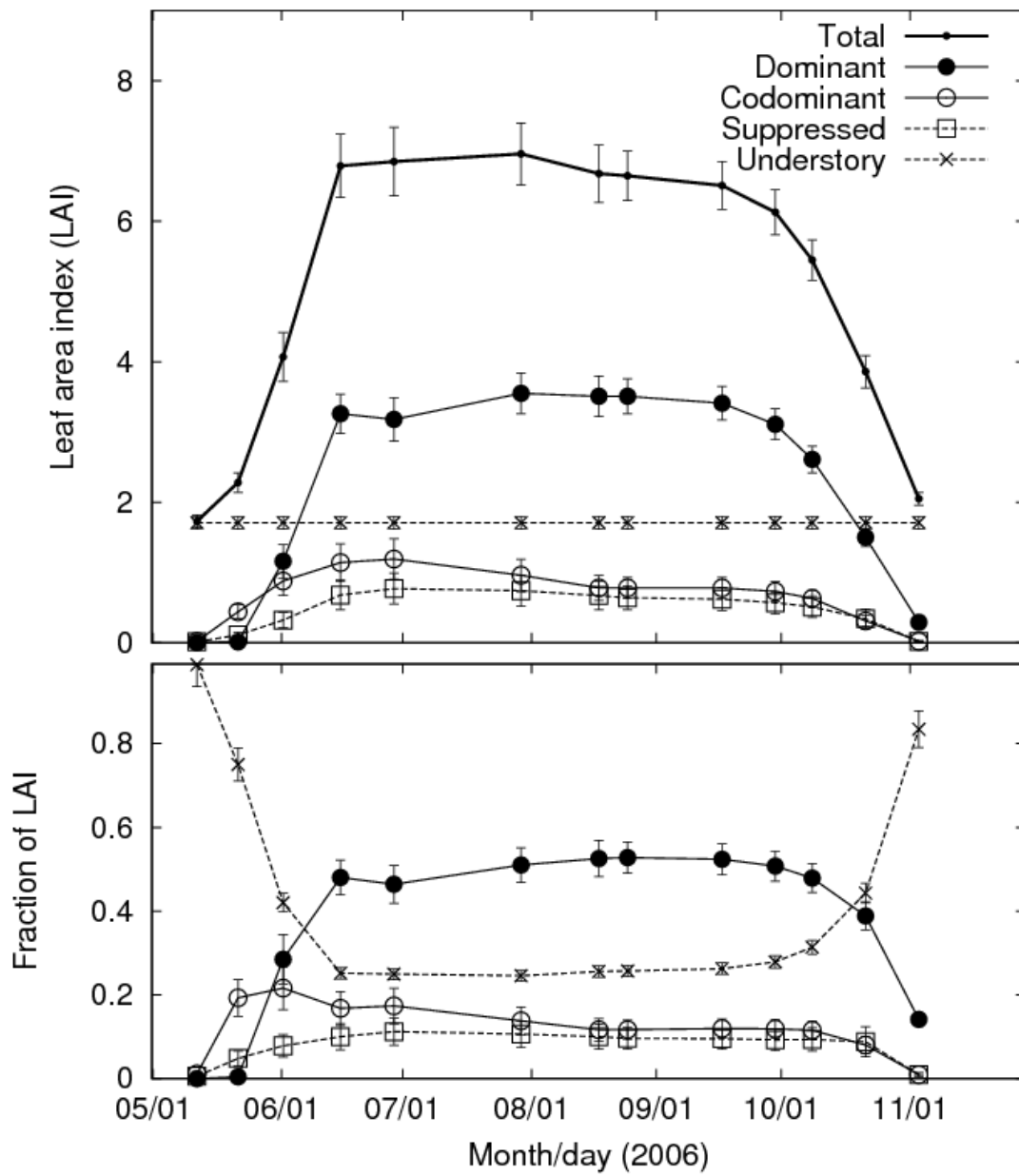


Fig. 3



751

Fig. 4

leaf area index (LAI) excluding Sasa.

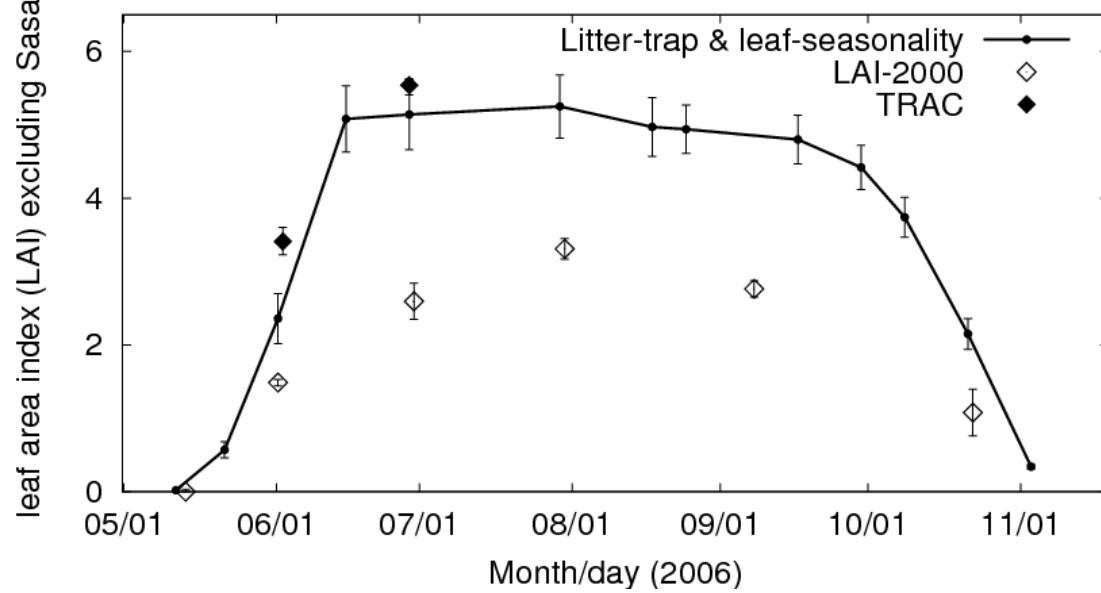
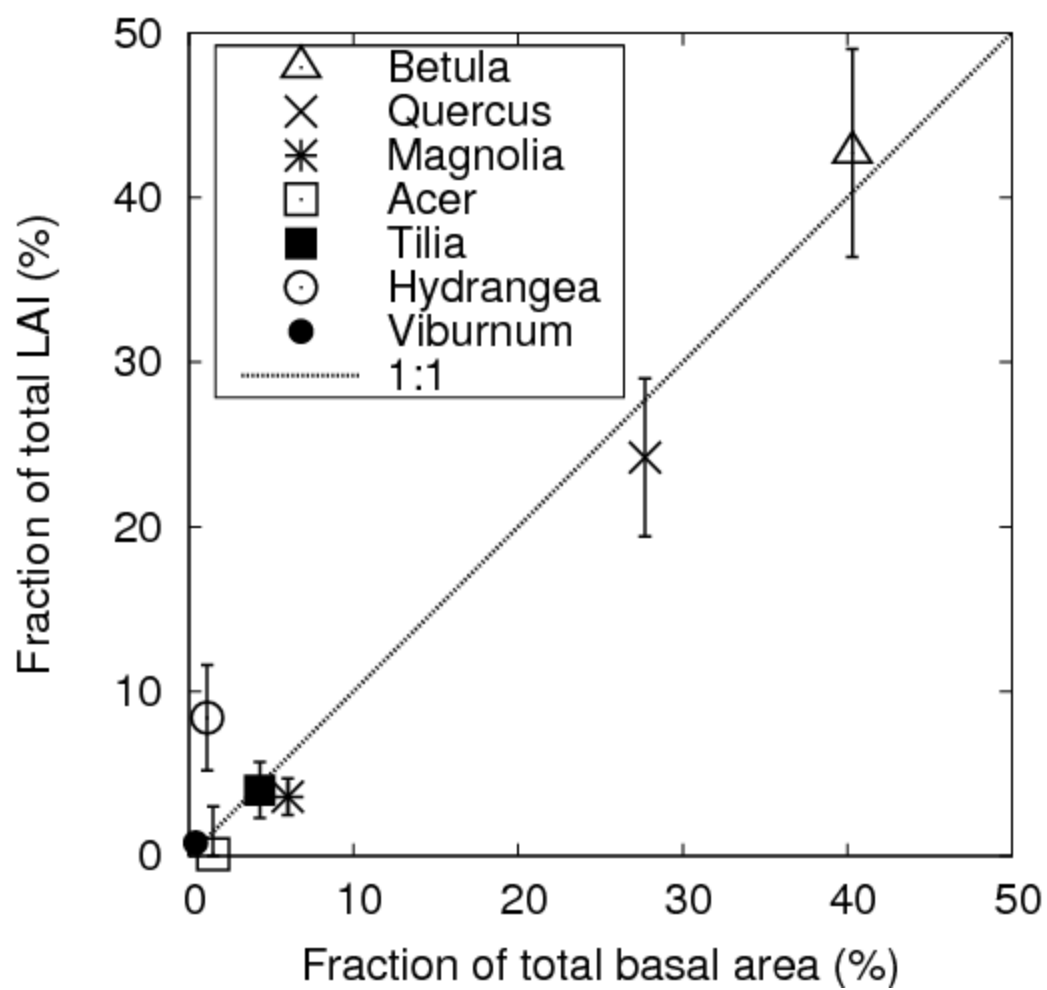


Fig. 5



756

Fig. 6